

"Carl Gogolak"
<cvlg@verizon.net>
02/15/2008 08:29 AM

To <Jay.Bassin@emsus.com>
cc
bcc

Subject RE: U.S. EPA Peer Review of Superfund PRG
Calculator

History:

Jay,

I have attached my comments as annotations to both the user's guide (yellow balloons) and the charge (blue insertion marks).

I looked at the references, but did not examine the parameters in detail. Generally, the User's Manual is not. It does not explain how to use the calculator, and does not describe the screen and input options. Without more info, tutorial or guidance, I would be reluctant to use anything but the defaults. As with many of these models, the interaction amongst the parameters is not spelled out, viz., if I change parameter A, should that imply a corresponding change to parameter B? Should certain parameters be restricted to a reasonable range? Should there be error checking on the input to prevent gross errors in their values?

There are MANY broken links: some crucial, like the equations in the calculator. Like many of the models out there, this one tries to include everything. There is no real discrimination between which factors are the biggest contributors and should be paid most attention to. There are sensitivity analyses that can be done to assess such things.

It may not simple or easy to make some of the changes/additions I suggest. I feel that most folks who are not experts in these particular models will just use the defaults.

Otherwise the learning curve may be too high.

I hope this is sufficient for the task at hand. If you need something else, I guess we have till COB today.

Regards

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From: Jay.Bassin@emsus.com [mailto:Jay.Bassin@emsus.com]

Sent: Wednesday, January 16, 2008 12:58 PM

To: carl@gogolak.org

Subject: U.S. EPA Peer Review of Superfund PRG Calculator

On behalf of EPA, I appreciate your willingness to conduct an external peer review of an on-line calculator of "Preliminary Remediation Goals for Radionuclides in Outdoor Surfaces."

We are requesting that you complete the peer review and provide me with your comments by February 15, 2008.

I am attaching a purchase order, the statement of work/"Charge" to the peer reviewers, a conflict of interest certification, and an IRS Form W-9 (the IRS form need not be used if your check will be issued to a corporation; if it will be issued to you as an individual, we are required to have it unless we already have yours on file).

Please return the conflict of interest certification as soon as convenient (not later than when you complete your peer review). Electronic submission is fine (either scanned signature or digital signature is acceptable).

Access instructions to the on-line calculator are provided in the SOW/Charge. Please contact me right away if you have problems accessing the site. We would appreciate it if you do not circulate the site, users name, or password.

While it is not germane to the peer review, I would appreciate it if you would keep track of how many hours you put into this peer review and let me know when you complete and return the purchase order for payment. This will assist us in designing future peer reviews.

Don't hesitate to call or email me if you have questions. Once again, I very much appreciate your willingness to help make this tool a better product.

(See attached file: Peer Review PO Gogolak.pdf)

(See attached file: SPRG Peer Review Charge 01-16-08.pdf)

(See attached file: Peer Reviewer Conflict of Interest Certification.pdf)

(See attached file: IRS Form w-9.pdf)

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www.emsus.com [users_guide_01142007 with cvg annotations.pdf](#)



Waste and Cleanup Risk Assessment

You are here: EPA Home OSWER Waste and Cleanup Risk Assessment Databases and Tools Remedial and Outdoor Surfaces SPRG Calculator

[http://risk.sprg.epa.gov/sprg/](#)
Last updated on Tuesday, January 13, 2009.

- SPRG Home
- SPRG Search
- User's Guide
- What's New
- Frequently Asked Questions
- Equations
- Download Area

Help for Key Concepts in Public Comments and System

Preliminary Remediation Goals for Radionuclides in Outdoor Surfaces (SPRG)

User's Guide

ABSTRACT: N/A

Disclaimer

This guidance document sets forth recommended approaches based on EPA's best thinking to date with respect to risk assessment for response actions at CERCLA sites. This document does not establish binding rules. Alternative approaches for risk assessment may be found to be more appropriate at individual sites. EPA encourages users to use an alternative approach and a description of any such approach should be placed in the Administrative Record for the site. Accordingly, if comments are received at individual sites questioning the use of the approaches recommended in this guidance, the comments should be considered and an explanation provided for the selected approach.

The policies set out in the radionuclide SPRG User Guide provide guidance to EPA staff. The User's Guide also provides recommended guidance to the public and regulated community on how EPA intends the National Oil and Hazardous Substances Pollution Contingency Plan (NCP) to be implemented. EPA may change this recommended guidance in the future, as appropriate. This calculator is intended for use by risk assessors, health physicists and other qualified environmental protection specialists.

It should also be noted that calculating a human radiological SPRG addresses neither human exposure risk from nonradiological (chemical) contaminants nor cancer toxicity, nor potential ecological risk. Of the radionuclides generally found at CERCLA sites, only uranium has potentially significant noncancer toxicity. When assessing sites with radiological contaminants which include uranium, it may also be necessary to consider the noncancer toxicity of uranium. Similarly, some sites with radiological contaminants in sensitive biological settings may also need to be evaluated for potential ecological risk. EPA's guidance "Ecological Risk Assessment Guidance for Superfund Sites: Designing and Conducting Ecological Risk Assessment" (EPA/600/R-03/001) provides information on the design and implementation of radiological risk assessments in sensitive biological settings. This guidance is available on the EPA's SPRG calculator.

This web calculator is intended to be a generic steady-state screening assessment tool. The calculator is flexible and may also be used to derive site-specific risk assessments. Site-specific information should be gathered. The use of models reviewed by EPA in the Soil Screening Guidance - Radionucleides Technical Background Document (EPA/600/R-03/001) is intended to be used for screening purposes. The calculator is not intended to be used for the derivation of site-specific Fate and Transport (PDF 383K, 25 pages) of the Soil Screening Guidance for Radionuclides. This report supports the information provided in Part 3 - Unsubstantiated Zone Models for Site-Specific Models to subsurface conditions and includes an assessment of each model's potential applicability to the soil screening process.

1. Introduction

Generally, these recommended radionuclide outdoor surface preliminary remediation goals (SPRGs) are reasonable maximum exposure (RME) concentrations derived from standardized equations that combine exposure information and toxicity information in the form of slope factors (SF). Recommended SPRGs are presented for residential and worker exposure.

The intent of this calculator is to address hard outside surfaces such as building sabs, outside building walls, sidewalks and roads.

Generally under the NCP, PRGs are risk-based, conservative screening values to identify areas and contaminants of potential concern (COPCs) that either do or do not warrant further investigation.

This calculator is based on Risk Assessment Guidance for Superfund, Volume I, Human Health Evaluation Manual (Part B, Development of Risk-based Preliminary Remediation Goals) (EPA/600/R-91/010). The calculator is based on the standard risk-based SPRGs that are modified based on site-specific data gathered during the RI/FS study. SPRG development and screening should assist staff in streamlining the consideration of remedial alternatives. Chemical-specific SPRGs typically are from two general sources: (1) concentrations based on potential Applicable or Relevant and Appropriate Requirements (ARARs) and (2) risk-based concentrations. ARARs include concentration limits set by other environmental regulations such as Safe Drinking Water Act maximum contaminant levels (MCLs). A SPRG, and the focus of this database tool, can be risk-based calculations that set concentration limits using carcinogenic toxicity values under specific exposure conditions.

The recommended approach for developing remediation goals is to identify SPRGs at scoping, modify them as needed at the end of the RI or during the FS based on site-specific information from the baseline risk assessment, and ultimately select remediation levels in the Record of Decision (ROD). In order to set radionuclide-specific remediation goals, the calculator uses the standard risk-based SPRG equations. The calculator also allows for the use of site-specific radionuclide-specific SPRGs. The recommended SPRG calculator provides the ability to modify the standard default SPRG exposure parameters to calculate site-specific SPRGs.

This database tool presents recommended standardized risk-based SPRGs and variable risk-based SPRG calculation equations for radioactive contaminants. Ecological effects are not considered in the calculator for radionuclides SPRGs.

SPRGs are presented for residents and workers using both soil volume and ground plane slope factors. The recommended risk-based SPRGs for radionuclides are based on the carcinogenicity of the contaminants. Cancer slope factors used are from HEAST.

Non-carcinogenic effects generally are not considered for radionuclide analytes, except for uranium for which both carcinogenic and non-carcinogenic effects are considered.

The standardized recommended SPRGs are based on default exposure parameters and incorporate site exposure factors that present RME conditions. This database tool also allows for the use of site-specific exposure parameters. Once this database is used to derive recommended standard SPRGs or calculate site-specific SPRGs, it is important to clearly demonstrate the calculations used to derive the SPRGs. The recommended SPRG calculator provides the ability to generate a Remedial Investigation (RI) Report or Feasibility Study. The calculations used to derive the SPRGs are presented, such as a Remedial Investigation (RI) Report or Feasibility Study.

This website combines current EPA recommended slope factors with "standard" exposure factors to estimate contaminant concentrations in environmental media (hard surfaces and soil) and the risk of humans (individual sensitive groups) over a lifetime. Significant knowledge about a given site may warrant the use of site-specific exposure parameters which differ from the default values. The recommended SPRG concentrations presented on this website can be used to screen pollutants in environmental media, trigger further risks is appropriate. The recommended SPRG concentrations presented on this website can be used to screen pollutants in environmental media, trigger further

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Refer reader to where it is explained how to do this.

Investigation, and provide initial cleanup goals, if applicable. The recommended SPRGs should be applied in accordance with guidance from EPA Regions.

2. Understanding the SPRG Website

2.1 General Considerations

Generally, these recommended SPRGs are based on concentration that correspond to specific levels of risk in dirt, smokes, sediments, finite soils and building materials. Slope factors (SFs), for a given radionuclide generally represent the risk equivalent per unit level (i.e., the risk equivalent per unit level of the radionuclide). In risk assessments, these SFs normally are used in calculations with radionuclide concentrations and exposure assumptions to estimate cancer risk from exposure to radionuclide contamination. The calculations may be rearranged to generate SPRGs for a specified level of risk. SFs may be specified for specific body organs or tissues of interest or for a weighted sum of individual organ doses. **Generic recommended SPRGs** are provided for each radionuclide and are multiplied by the total activity of each radionuclide inhaled or ingested to estimate the cancer risk to the receptor. Cancer slope factors from HEST and FS13 are used to estimate cancer risk.

The most common land use and exposure assumptions are included in the equations on this website: Residential, Outdoor Worker, and Indoor Worker. The recommended SPRGs are generated with standard exposure, intake equation  EPA site-specific exposure parameters.

2.2 Slope Factors (SFs)

EPA classifies all radionuclides as carcinogenic to humans. The radionuclide table from HEST lists ingestion, inhalation and external exposure cancer slope factors (risk coefficients for total cancer morbidity) for radionuclides in conventional units of picocuries (pCi). Ingestion and inhalation slope factors are central estimates in a linear model of cancer risk. External exposure slope factors are central estimates of lifetime attributable cancer risk per unit of activity inhaled or ingested, expressed as risk/pCi. External exposure slope factors are central estimates of lifetime attributable cancer risk per unit of activity inhaled or ingested, expressed as risk/pCi. External exposure slope factors are distributed uniformly in a thick layer of soil, and are expressed as risk/yr per pCi/gm soil. When combined with site-specific radionuclide data and appropriate exposure assumptions, slope factors can be used to estimate lifetime cancer risk to members of the general population due to radionuclide exposures.

2.2.1 When To Use "+D" SPRGs

Several of the isotopes are listed with a "+D" designation. This designation indicates that the SF includes the contribution from ingrowth of daughter isotopes out to 100 years. The intention of this designation is to make realistic PRGs by including the contributions from their short-lived decay products, assuming equal activity concentrations (i.e., secular equilibrium) with the parent or parent nuclide in the environment. (Note that there is one exception to the assumption of equal activity concentrations: Pu-239 and Pu-240 are listed with "+D" designations, but their SFs are based on the assumption of secular equilibrium with their parent, Pu-238, and not Pu-239 and Pu-240 in air.) Before applying PRGs to a site, it should be determined if the radionuclides are in secular equilibrium with their parent. If the radionuclides are in secular equilibrium, the +D PRGs should be used for the parent isotope and the daughters included in the +D can be ignored. If the isotopes are not in secular equilibrium, the +D PRGs should be applied for each daughter isotope. However, in the absence of empirical data, the "+D" values for radionuclides should be used unless there are compelling reasons not to.

For example, if analytical data from a site reveal that Th-232, Ra-224, Ra-226, Rn-220 are detected at a site and that they are in secular equilibrium, the PRG for Th-232+D should be applied and the Ra-224 and Rn-220 can be ignored.

Another example would concern a decay chain in secular equilibrium like Th-232. Even though the decay chain for Th-232 is very long, there is no Th-232+D slope factor. For this chain, the PRG for Th-232 is based on the assumption of secular equilibrium. The user should use each of the PRGs for the other radionuclides in the decay chain that are in secular equilibrium. For example, if Ra-226, Ra-228, Th-230, Th-232, Ra-224, Ra-228, Ac-228, Th-230, Bi-210, Bi-212, Po-210, Po-212, Pb-210, Pb-212, and Pb-214 are detected at a site and that they are in secular equilibrium, then the user may use that particular +D slope factor that covers that part of the decay chain. While using the slope factors for the other radionuclides.


2.2.2 Associated Decay Chains for "+D" SPRGs

Selected radionuclides and radioactive decay chain products are designated with the suffix, "+D" to indicate that cancer risk estimates for these radionuclides include the contributions from their short-lived decay products, assuming equal activity concentrations (i.e., secular equilibrium) with the parent or parent nuclide in the environment. For all radionuclides without the "+D" suffix, only intake or external exposure to the single radionuclide is considered. Most radionuclides with a +D year. This table provides the associated decay chain included with the terminal radionuclide used in the slope factors. This table is reproduced below.

Principal Radionuclide (half-life in years)	Associated decay chain	Terminal Radionuclide (years)	Half-life (years)
Am-241m-D (152)	Am-242, Cm-242, Np-238	Pu-238	87.7
Am-243+D (7.4E+03)	Np-239	Pu-239	2.40E+04
Np-237+D (2.1E+06)	Pu-233	U-233	1.6E+05
Pu-244+D (8.3E+07)	U-240, Np-240m	Pu-240	6.50E+03
Ra-226+D (1.6E+03)	Rn-222, Po-218, Pb-214, Ac-218, Bi-214, Po-214, Tl-210	Pb-214	22
Ra-228+D (6)	Ac-228	Th-228	2
U-235+D (7.0E+08)	Th-232	Pb-231	3.3E+04
U-238+D (4.5E+09)	Th-234, Pa-234m, Pb-234	U-234	2.4E+05

Ingestion and inhalation slope factors are missing for some of the +D isotopes. These have not been derived yet. Use caution when selecting a SPRG to make sure that as many routes of exposure are accounted for.

2.3 SPRG in Context of Superfund Modeling Framework

This recommended SPRG calculator focuses on the application of generic and simple site-specific approaches that are part of a larger framework for calculating concentration levels that are designed to be consistent with risk based criteria. Generic recommended SPRGs for a 1×10^{-6} cancer risk standard are provided by viewing the calculator. The calculator is available by clicking on the SPRG icon  of this calculator with the Calc Default SPRGs option. The 3 of the Soil Screening Guidance for Risk-Based Technical Backscatter Assessment provides more information about more detailed approaches that are part of the same framework.

Generic recommended SPRGs can be calculated from the same equations presented in the site-specific portion of the calculator, but typically are based on a number of assumptions. In general, they are based on a human health for most site conditions. Generic recommended SPRGs can be used in place of site-specific SPRG levels; however, in general, they are intended to be used as a screening tool. The calculator provides the cost of collecting the data necessary to develop site-specific SPRGs with the potential for deriving a higher SPRG that provides an appropriate level of protection.

https://epa-sprg.coml.gov/users_guide.shtml

1/15/2008

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Author: Carl Gogolak
Subject: None
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Author: Carl Gogolak
Subject: None
Date: 2/14/2008 4:14:10 PM
Many of these links say authorization required. Why is this since I already entered a password?

3. Using the SPRG Table

The SPRG "Download Area" provides generic recommended concentrations in the absence of site-specific exposure assessments. Screening concentrations can be used for:

- Protecting multiple sites within a facility or exposure units
- Setting risk-based detection limits for contaminants of potential concern (COPCs)
- Focusing future risk assessment efforts
- When appropriate for the site, consideration as risk-based cleanup levels

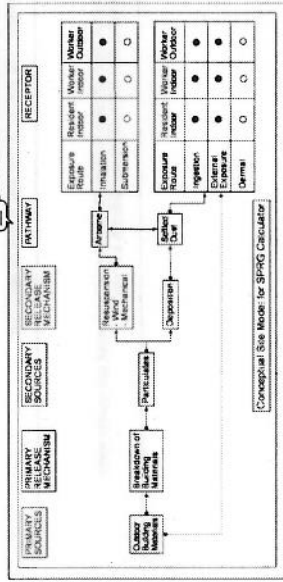
3.1 Developing a Conceptual Site Model

When using SPRGs, the exposure pathways of concern and site conditions should match those taken into account by the screening levels. (Note, however, that future users may not match current uses. Future uses of a site should be logical as conditions which might occur at the site in the future.) Thus, it generally is necessary to develop a conceptual site model (CSM) to identify likely dominant site use, exposure pathways and potential receptors. This information can be used to develop a CERCLA site model (CSM) to identify likely dominant site use, exposure pathways and potential receptors. This information can be used to develop a CERCLA site model (CSM) for ecological receptors, and routes and receptors based on historical information. It summarizes the understanding of the contamination problem. A separate CSM for ecological receptors can be useful. Part 2 and Attachment A of the Soil Screening Guidance for Radionuclides: Users Guide (EPA 2000a) contains the recommended steps for developing a CSM.

Existing EPA documents with additional CSM guidance are:

1. Risk Assessment Guidance for Superfund: Volume I Human Health Evaluation Manual (Part D, Standardized Planning, Reporting, and Review of Superfund Risk Assessments). See Planning Table 1; and
2. Soil Screening Guidance for Radionuclides: User's Guide. See Attachment A.

CSMs can be tabular, graphical or stem-and-leaf. Section 4.2.1 of the Users Guide presents links to graphical CSMs for each scenario. Below is a stem-and-leaf CSM showing the exposure routes quantified and not quantified in this calculator.



As a final check, the CSM generally should answer the following questions:

- Are there potential ecological receptors?
- Is there potential for land use other than those listed in the SPRG calculator (e.g., recreational, agricultural or trespasser)?
- Are there other likely human exposure pathways that were not considered in development of the SPRGs?
- Are there unusual site conditions (e.g., large areas of contamination, high fugitive dust levels, potential for indoor air contamination)?

The SPRGs may need to be adjusted to reflect the answers to these questions, and additional tools or assessment methodologies may need to be considered (e.g., if there may be potentially significant ecological risk). The recommended default scenarios in this calculator are the same default scenarios EPA addresses in its guidance. Other scenarios may be investigated, using the recommended SPRG calculator, by altering site-specific exposure parameters.

3.2 Radionuclide Background

Natural background radiation should be considered prior to applying SPRGs as cleanup levels. Background site-related levels of radiation should be addressed as they are for other contaminants at CERCLA sites. For further information see EPA's guidance "Rate of Background in the CERCLA Cleanup Program", April 2002. (OSWER 9285.6-07P). It should be noted that certain ARARs specifically address how to factor background into cleanup levels. For example, some radionuclide ARAR levels are established as increments above background concentrations. In these circumstances, background should be addressed in the manner prescribed by the ARAR. Additional information on radionuclide materials present in building materials can be found in Volume 105, Number 2, March/April 2000, Journal of Research of the National Institute of Standards and Technology, "Radionuclide Measurements on Selected Domestic Surfaces."

3.3 Potential Problems

As with any risk based tool, the potential exists for misapplication. In most cases, this results from not understanding the intended use of the recommended SPRGs. In order to prevent misuse of the recommended SPRGs, the following should be avoided:

- Applying recommended SPRG levels to a site without adequately developing a conceptual site model that identifies relevant exposure pathways and exposure scenarios.
- Using recommended SPRG levels as cleanup levels without the consideration of other relevant criteria such as ARARs.
- Use of outdated SPRG tables that have been superseded by more recent publications.
- Not considering the effects from the presence of multiple isotopes.

4. Technical Support Documentation

The recommended SPRGs consider human exposure from direct contact with contaminated outdoor dust on solid surfaces and external exposure to contaminated streets, sidewalks, finite slabs and building materials. The equations and technical discussion are aimed at developing concentration levels for risk-based SPRGs. The following text presents the recommended land use equations and their exposure routes. Table 1 presents the suggested definitions of the variables and their default values. Any alternative values or assumptions used in remedy evaluation or selection on a CERCLA site should be presented with supporting rationale in the Administrative Record.

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Author: Carl Gogolak
Subject: Note
Date: 2/14/2008 4:38:08 PM
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Author: Carl Gogolak
Subject: Note
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Author: Carl Gogolak
Subject: Note
Date: 2/15/2008 7:48:21 AM
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Author: Carl Gogolak
Subject: Note
Date: 2/15/2008 7:58:05 AM
I guess open circles signify not quantified?

Author: Carl Gogolak
Subject: Note
Date: 2/15/2008 7:49:07 AM
In the Surface PRG calculations, the 3 choices are resident, indoor worker and outdoor worker. Where do these other scenarios fit in?

$$SPR_{2-D} = \frac{C_{soil} \cdot V_{soil} \cdot SF}{C_{air} \cdot V_{air} \cdot SF} \cdot \frac{1}{(25 \text{ years})^{1.4}} \cdot \left(\frac{1}{24 \text{ hours}} \right)$$

$$C_{soil} = \frac{C_{air} \cdot V_{air} \cdot SF}{V_{soil} \cdot SF} \cdot \left(\frac{250 \text{ days}}{\text{year}} \right) \cdot \left(\frac{1 \text{ day}}{24 \text{ hours}} \right)$$

The resulting units for this recommended SPRG are in pCi/g. The units are based on mass because the SF used is the soil volume for external exposure.

• 2-D Exposure to Direct External Exposure (Materials with fixed contamination in a finite slab at 15cm depth using 15cm soil volume toxicity values)

$$SPR_{2-D} = \frac{C_{soil} \cdot V_{soil} \cdot SF}{C_{air} \cdot V_{air} \cdot SF} \cdot \frac{1}{(25 \text{ years})^{1.4}} \cdot \left(\frac{1}{24 \text{ hours}} \right)$$

$$C_{soil} = \frac{C_{air} \cdot V_{air} \cdot SF}{V_{soil} \cdot SF} \cdot \left(\frac{250 \text{ days}}{\text{year}} \right) \cdot \left(\frac{1 \text{ day}}{24 \text{ hours}} \right)$$

The resulting units for this recommended SPRG are in pCi/g. The units are based on mass because the SF used is the soil volume for external exposure.

• 2-D Exposure to Direct External Exposure (Materials with fixed contamination in a finite slab using ground glass toxicity values)

$$SPR_{2-D} = \frac{C_{soil} \cdot V_{soil} \cdot SF}{C_{air} \cdot V_{air} \cdot SF} \cdot \frac{1}{(25 \text{ years})^{1.4}} \cdot \left(\frac{1}{24 \text{ hours}} \right)$$

$$C_{soil} = \frac{C_{air} \cdot V_{air} \cdot SF}{V_{soil} \cdot SF} \cdot \left(\frac{250 \text{ days}}{\text{year}} \right) \cdot \left(\frac{1 \text{ day}}{24 \text{ hours}} \right)$$

The resulting units for this recommended SPRG are in pCi/cm². The units are based on area because the SF used is the ground glass for external exposure.

4.3 Exposure Parameter Justification

The following sections describe the exposure parameter default variables and the values selected.

4.3.1 Exposure Time (ET)

The exposure time represents the hours per day that a receptor spends exposed to a source. The exposure times vary by exposure scenario, age of the receptor and the location of the exposure. For children, the exposure time is based on the EPA Office of Pesticide Programs (OPP) For Inhalation and External Exposure the exposure time indoors is set at 16.4 hours per day and the exposure time outdoors is set at 1,752 hours per day. These values are from the 1997 Exposure Factors Handbook. Note, that inhalation and subsequent ingestion of dust particles trapped in mucous is not quantified in the SPRG equations due to lack of exposure information.

For the outdoor and indoor worker, exposure time for the dust ingestion exposure route is based on exposure to hard surfaces. For this calculator, the defaults were set at 8 hr/d. The exposure time for direct external exposure is the entire work day or 8 hr/d.

4.3.2 Fraction Transferred from Surface to Skin (FTSS)

In general, this is the fraction of residue on a surface that can be transferred to skin. US EPA 2003 (pp D-5) states that hand press experiments were conducted on dry skin. Transfers of 50% were observed for hard surfaces. These are considered representative of the WTC situation and were adopted for this calculator.

4.3.3 Surface Area (SA)

In general, this is the skin area contacted during the mouthing event. The OPP default is 20 cm² based on the surface area of the 3 fingers that a child will most likely use for hand to mouth transfer. Total skin surface area increases by about 3 fold from age 2 to an adult. Average area of both hands for an adult is about 900 cm², so it would be about 300 cm² for a 2 year old. Assuming 3 fingers of one hand represents about 5% of the total area of both hands, it would increase from 15 cm² to 45 cm² from age 2 to adult. On this basis, the SA values used here are assumed to start at 15 cm² and increase linearly to 45 cm² at age 17 and remain constant after that.

4.3.4 Frequency of Hand to Mouth (FH)

The OPP defaults suggest 9.5 events/hr for toddlers, based on observations at day care centers. This will decline with age, but very little data are available for other ages. Michael et al (1994) assumed a mouthing frequency of twice per day for adults. It was decided to group the age cohort-specific hand-to-mouth frequency as follows: 1 to 6 yr - 9.5 times/hr, 7 to 12yr - 5 times/hr, 8 to 18 yr - 2 times/hr and 19 to 31 yr - 1 time/hr.

4.3.5 Salivary Extraction Factor (SE)

In general, the fraction transferred from skin to mouth will depend on the contaminant, mouthing time and other behavioral patterns. The OPP default is 50%, based on pesticide studies. Michael et al (1994) assumed that all of the residues deposited on the fingertips would be transferred to the mouth, twice per day. In the Binghamton re-entry guideline derivation, a range of factors were used: 0.05, 0.1, and 0.25 representing the fraction of residue on hand that is transferred to the mouth (Kim and Hanley, 1993). For purposes of this assessment, the OPP default of 50% was selected for all ages.

4.3.6 Resident Age-Adjusted Dust Ingestion Rate (IR)

To account for the variability in exposure activities between children and adults, the age-adjusted dust ingestion rate equation was developed. This equation takes into account the differences in hand to mouth behavior, hand surface area, and exposure to hard and soft surfaces over the exposure durations of an adult and child.

4.3.7 Worker Dust Ingestion Rate (IR_w)

This dust ingestion equation calculates the intake for a worker based on exposure to hard surfaces.

4.3.8 Dispersation Rate Constant (k)

In some circumstances, the load of dust on a contaminated surface, to which receptors are exposed, may decline over time. Dispersation of dust may result from weather, cleaning and transfer to skin and clothing. Different surfaces may be cleaned at different rates and any dispersation rate used should consider a representative cleaning frequency. To determine whether dispersation is a factor at a given site, the site manager should establish whether a significant reservoir of contaminated dust is present. Such reservoirs may function as sources of dust and negate the impacts of dispersation mechanisms. The first step in identifying the presence of a reservoir is to examine the site history. If a reservoir is identified, the site manager should consider the impacts of dispersation on the exposure assessment. When reservoirs are less likely to exist, such as at sites where contamination is the result of a single spill, dust cloud or event, it may be more important to account for dispersation of surface loads. For fixed contamination in materials (outside walls, streets and pavement), or on material surfaces, in the 3-D and 2-D equations, the dispersation term is not included as dispersation is not expected.

The recommended default value for the dispersation rate constant is 0.0. This assumes that a contaminant reservoir is present. However, the variable is adjustable in the SPRG calculator. If a dispersation rate constant is used, it is assumed that the dust was deposited as a one time event (i.e.: dust cloud). Also, if a dispersation rate is applied, it is assumed that it is applicable from the point in time the SPRG is calculated into the future. The discussion below provides a review of the indoor surfaces

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Author: Carl Gogolak

Date: 2/14/2008 5:23:16 PM

Subject: Note
Can these discussions be hyperlinked to Table 1? There is a lot of info here, but it is not very well organized to be useful. Since this appears to be a web tool, maximum advantage of hyperlinking should be considered. It would also be useful to have a self contained help file with the same links.

Author: Carl Gogolak

Date: 2/15/2008 2:33:58 AM

Subject: Note
Did I miss any prior discussion of how this tool should be used differently for children and adults?
Since this is screening, would using the most restrictive be appropriate?, for residential?

roadway classes. Simple vehicle navigation instructions are available. To quickly get to the functional class information make sure the "Highway Information" tab is selected and then make sure the drop-down menu under "General Map" indicates "Functional Class". Now the user can use the zoom controls to reach the area of interest. This resource could be considered to apply site-specific inputs for calculating ADTV and SIF for a risk assessment. Further state-specific information can be found by contacting the contractor.

4.3.11 Area Correction Factor

The RAC3/HHEM Part B model assumes that an individual is exposed to a source geometry that is effectively an infinite slab. The concept of an infinite slab means that the volume of the contaminant zone and its serial extent are so large that it behaves as if it were infinite in its physical dimensions. In practice, soil contaminated to a depth greater than about 15 cm and with an aerial extent greater than about 1,000 m² will create a radiation field comparable to that of an infinite slab. (U.S. EPA, 2006a)

To accommodate the fact that in most residential settings the assumption of an infinite slab source will result in overly conservative SS_{2,4} an adjustment for source area geometry is incorporated into the RAC3/HHEM Part B model. This, an area correction factor, ACF, has been added to the calculation of recommended exposure models addressing outside walls, streets and sidewalks. For the 2-D exposure model, the ACF is presented for all isotopes for the 2-D site-specific analysis. This calculator allows the user to select from 8 different slab area sizes. If no size is selected for the finite slab analysis, the ACF from the most protective slab size is selected. For further information on the calculation of the site-specific/area specific ACF values for 2-D slabs see Contaminated Slabs for a description of other EPA Default ACF values, follow the link here [\[Link\]](#)

4.3.13 Surfaces Factor

The 3-D direct external exposure equations (Building materials and dust) without F_{air} are single surface equations. The surfaces factor, in the default and site-specific equations, are based on exposure to 2 vertical surfaces (outside building surfaces on either side of a street) and a horizontal surface (road and sidewalk). This calculator surfaces factor (F_{sur}). The default surface coefficients for exposure in a contaminated outdoor setting and dose rate coefficients for an infinite source to calculate a surfaces factor (F_{sur}) are based on the geometry of the building and the location of the exposure receptor. The 3-D exposure model is modified to account for the dose contribution from multiple surfaces. Further, photon energies of each radionuclide were incorporated into the modeling. Please see the attached pop file for detailed explanation of the process. Site Walk Dose Rate shows that building height doesn't effect the dose rate significantly after 150 feet. The surfaces factor is a table of the F_{sur} values used in this calculator for each radionuclide. F_{sur} values were calculated for each position-specific and building height specific condition.

4.4 Supporting Equations

There are two parts in the above and use equations that require further explanation. First is the use of the radionuclide decay constant (λ). Second is the variable particulate emission factor feature of this calculator.

4.4.1 Radionuclide Decay Constant

Each equation (where appropriate by media) has a decay constant term which is based on the half-life of the isotope (T_{1/2}). λ = Decay constant (0.693/half-life in years). The location of this term is to derive radiologic specific for radionuclides with relatively short half-lives, compared to the exposure duration (ED). The term $(1 - e^{-\lambda})$ takes into account the number of half-lives that will occur within the ED to calculate an appropriate value. Derivations of the input variables are in Tables 1.

4.4.2 Particulate Emission Factor (PEF)

Two particulate emission factors can be selected for this calculator: mechanically driven and the traditional wind driven emission factor.

4.4.2.1 Mechanically Driven PEF

This equation allows the user to input vehicle weight, road dimensions, distance traveled and time. Although developed for unpaved roads, this equation can be used to estimate emission factors after an incident. The receptor is assumed to be exposed to contaminants in the form of particulate matter with an aerodynamic particle blowing on surfaces. PEF for wind driven emission factors is based on the relationship between wind speed and the amount of dust emitted from the road. PEF for wind driven emission sources. For this reason, the PEF for unpaved road traffic and the PEF for wind erosion are calculated separately.

The following fugitive dust emission equation represents approximation of actual emissions at a specific site. Sensitive emission model parameters include the soil lift coefficient (K_s), dust emission factor (EF_d), dust emission factor (EF_w) in diameter and can be measured as the proportion of soil passing a 200-mesh standard mesh screen. EF_d is defined as soil particles smaller than 75 micrometers (µm) in diameter and can be measured as the proportion of soil passing a 200-mesh standard mesh screen. EF_w is defined as soil particles larger than 75 micrometers (µm) in diameter and can be measured as the proportion of soil passing a 100-mesh standard mesh screen. The dust emission factor (EF_d) is defined as the amount of dust emitted per unit area of road surface per unit time. In general, soil lift is dependent on wind speed, soil moisture content, soil texture, and soil structure. Model parameters for which default values have been assigned, however, site-specific values will produce more accurate modeling results. Other emission model parameters have not been assigned default values and are typically defined on a site-specific basis. These parameters include the total distance traveled by vehicles, mean vehicle weight, average vehicle speed, and the area of roadway.

Mean vehicle weight (W) in tons is calculated by determining vehicle weight classes and numbers in that class. An example is presented below for site specific data. The default mean vehicle weight selected for this calculator is 3.2 tons based on page 4-285 in PROCEDURES DOCUMENT FOR NATIONAL EMISSION INVENTORY CRITERIA AIR POLLUTANTS 1995-1999. EPA-454/R-01-006. However, there is wide variation in vehicle weights when considering industrial facilities. AR42 supporting documentation reveals in Table A1-5 that the mean vehicle weight can range up to 42 tons. Site-specific conditions should be considered or measured. The table is reproduced below.

Author: Carl Gogolik
 Subject: None
 Date: 2/15/2008 2:35:49 AM
 Link refers to Table 5.1.
 Is there supposed to be a Section 5 in this user's manual?

GSF _I	Gamma Shielding Factor - Indoor (unitless)	0.4 (assumes shielding)	EPA 2000b. (pg. 2-18) Other GSFs are presented in these reports. U.S. EPA 2000a. (pg. 2-22), U.S. EPA 2000b. (pg. 2-18)
F _{AM}	Area and Material Factor (unitless)	1.0	ANL 2001 (Fig 8.6)
F _{CD}	Depth and Cover Function (unitless)	1.0	ANL 2001 (Fig 8.6)
F _{OFF-SET}	Off-set Factor (unitless)	1.0	ANL 2001 (Fig 8.6)
F _{SURF}	Surfaces Factor (unitless)	Isotope-specific	Eckerman, 2007
Inhalation and Ingestion Rates			
IF _w	Worker Dust Ingestion Rate - Worker (cm ³ /day)	90	Calculated Value based on EPA 2003 (pg. D-4)
IF _r	Age-Adjusted Dust Ingestion Rate - Resident (cm ³ /day)	64.5	Calculated Value based on EPA 2003 (pg. D-4)
HR _r	Age adjusted Inhalation Rate (m ³ /day)	18	Calculated Value based on U.S. EPA 1991 (pg. 15)
HR _a	Adult Inhalation Rate (m ³ /day; based on IRIS default)	20	U.S. EPA 1991 (pg. 15)
HR _c	Child Inhalation Rate (m ³ /day; based on IRIS default)	10	U.S. EPA 1991 (pg. 15)
HR _w	Worker Inhalation Rate (m ³ /hr)	2.5	U.S. EPA 1997 (pg. 5-25)
Ingestion Rate Variables			
FTSS _h	Fraction Transferred Surface to Skin - Hand Surface (unitless)	0.5	EPA 2004 (Exhibit E-1 pg. E-6)
SA _a	Surface Area of Fingers - Adult (cm ²)	45	EPA 2003 (pg. D-5)
SA _c	Surface Area of Fingers - Child (cm ²)	15	EPA 2003 (pg. D-5)
SA _w	Surface Area of Fingers - Worker (cm ²)	45	EPA 2003 (pg. D-5)
FQ _a	Frequency of Hand to Mouth - Adult (events/hour)	1	EPA 2003 (pg. D-5)
FQ _c	Frequency of Hand to Mouth - Child (events/hour)	9.5	EPA 2003 (pg. D-5)
FQ _w	Frequency of Hand to Mouth - Worker (events/hour)	1	EPA 2003 (pg. D-5)
SE	Saliva Extraction Factor (unitless)	0.5	EPA 2003 (pg. D-5)
ET _{h,a}	Exposure Time - Adult Hand Surface (hours/day)	4	EPA 2003 (pg. D-4)
ET _{h,c}	Exposure Time - Child Hand Surface (hours/day)	4	EPA 2003 (pg. D-4)
ET _{h,w}	Exposure Time - Worker Hand Surface (hours/day)	8	EPA 2003 (pg. D-4)
ET _{o,r}	Exposure Time Outdoor - Resident (hours/day)	1.752	EPA 1997 (Table 15-132)
ET _{i,r}	Exposure Time Indoor - Resident (hours/day)	16.4	EPA 1997 (Table 15-131)
ET _w	Air-Exposure Time - Worker (hours/day)	8	EPA 2003 (pg. D-4)
Exposure Frequency, Exposure Duration, and Exposure Time Variables			
EF _{iw}	Exposure Frequency - indoor worker (days/year)	250	U.S. EPA 1991 (pg. 15)
EF _{ow}	Exposure Frequency - outdoor worker (days/year)	225	U.S. EPA 1991 (pg. 15)
EF _r	Exposure Frequency - resident (days/year)	350	U.S. EPA 1991 (pg. 15)
ED _w	Exposure Duration - worker (years)	25	U.S. EPA 1991 (pg. 15)
ED _r	Exposure Duration - resident (years)	30	U.S. EPA 1991 (pg. 15)
ED _a	Exposure Duration - adult resident (years)	24	U.S. EPA 1991 (pg. 15)
ED _c	Exposure Duration - child resident (years)	6	U.S. EPA 1991 (pg. 15)
Particulate Emission Factor Variables			
PEF _w	Wind Particulate Emission Factor - Minneapolis (m ² /kg)	1.36 x 10 ⁸ Minneapolis-specific	U.S. EPA 1996a (pg. 23), U.S. EPA 1996b (pg. 31)
O/C _w	Inverse of the Mean Concentration at the Center of a 0.5-Acre-Square Source - wind(g/m ² -s per kg/m ³)	93.77 Minneapolis-specific	U.S. EPA 1996a (pg. 23), U.S. EPA 1996b (pg. 31)
V	(fraction of vegetative cover) unitless	0.5	U.S. EPA 1990b, U.S. EPA 1996a (pg. 23), U.S. EPA 1996b (pg. 31)
U _m	mean annual wind speed) m/s	4.69	U.S. EPA 1990b, U.S. EPA 1996a (pg. 23), U.S. EPA 1996b (pg. 31)
U _t	equivalent threshold value of wind speed at 7m)	11.32	U.S. EPA 1990b, U.S. EPA 1996a

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Subject: Note

Date: 2/14/2008 4:50:20 PM

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EDUCATION

B.S., Physics, Fordham University, Bronx NY, 1970

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Ph.D., Applied Statistics, Polytechnic University, Brooklyn NY, 1986

POSITIONS

2006 – present Consultant on Environmental Radiation Physics and Statistical Survey Design

1970 - 2005 Physicist, Applied Physics Division, USDHS/EML

Principal Investigator, Outreach/Educational Program

1994 - 1995 Acting Director, Radiation Physics Division, USDOE/EML

1990 - 1992 New Production Reactor Research Coordinator, USDOE/EML

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1987 - 1995 Supervisor, Radon Group, USDOE/EML

1986 - 1995 Adjunct Mathematics Lecturer, Polytechnic University

1982 - 2002 Guest Lecturer, Harvard University School of Public Health Short Course Program

1980, 1983, 1986 Visiting Scientist, Institute for Radiation Hygiene, Federal Health Office, Munich

Dr. Gogolak conducts experimental and theoretical studies of ionizing radiation required for preventing, protecting against, and responding to radiological and nuclear events. His earlier work included research on the impact of non-nuclear energy technologies on the environment and the effect of diurnal radon progeny concentrations on terrestrial gamma-ray exposure rates; testing real-time monitoring instrumentation for ⁸⁵Kr releases from nuclear reprocessing plants; developing the first mobile computer system for on-site data analysis of in-situ gamma-ray measurements and designing a monitoring program for potential releases from light water reactors employing high pressure ionization chambers and thermoluminescent dosimeters.

Dr. Gogolak was a member of a team which developed statistical procedures for the Multi-Agency Radiation Survey and Site Investigation Manual (MARSSIM). This team was awarded Vice President Albert Gore's Hammer Award for Reinventing Government. In addition, he developed statistical methods for linking data quality objectives and measurement quality objectives for the Multi-Agency Radiological Analytical Laboratory Protocols (MARLAP) Manual. He developed the kriging methodology used by the National Cancer Institute to determine the thyroid dose from ¹³¹I releases from atmospheric testing in Nevada during the 1950's and 1960's.

PROFESSIONAL ACTIVITIES

Dr. Gogolak is a member of the National Council on Radiation Protection and Measurements Scientific Committee 64-2, A Design of Effective Effluent and Environmental Monitoring Programs. He is a co-Chair of the Council on Ionizing Radiation Measurements and Standards subcommittee on Public/Environmental Radiation Protection.

He is a Fellow of the Health Physics Society

He has authored or co-authored more than 100 papers, presentations, and technical reports.

Peer Reviewer Conflict of Interest Certification

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(b) Peer Reviewer agrees that if an actual, apparent, or potential personal or organizational conflict of interest is identified during performance of this peer review, he/she immediately will make a full disclosure in writing to EMS. This disclosure shall include a description of actions that Peer Reviewer (or his/her employer) has taken or proposes to take after consultation with EMS to avoid, mitigate, or neutralize the actual, apparent, or potential organizational conflict of interest. Peer Reviewer shall continue performance until notified by EMS of any contrary action to be taken.

Signature: [Handwritten Signature] Date: 1/28/08

[] Check here if any explanation is attached

Printed Name: CARL V. GOGOLAK

Affiliation/Organization: EMS